

Allee effects and the evolution of polymorphism in cyclic parthenogens.

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Abstract

Cyclic parthenogens alternate asexual reproduction with periodic episodes of sexual reproduction. Sexually produced free-living forms are often their only way to survive unfavorable periods. When sexual reproduction requires the mating of two self-incompatible individuals, mating limitation may generate an Allee effect, which makes small populations particularly vulnerable to extinction; parthenogenetic reproduction can attenuate this effect. However, asexual reproduction likely trades off with sexual reproduction. To explore the evolutionary implications of such a trade-off, we included recurrent mating events associated with seasonal interruptions in a simple population dynamics model. Following an adaptive dynamics approach, we showed that positive density dependence associated with Allee effects in cyclic parthenogens promotes evolutionary divergence in the level of investment in asexual reproduction. Although polymorphism may be transient, morphs mostly investing into sexual reproduction may eventually exclude those predominantly reproducing in an asexual manner. Asexual morphs can be seen as making cooperative investments into the common pool of mates, while sexual morphs defect, survive better, and may eventually fix in the population. Our findings provide a novel hypothesis for the frequent coexistence of sexual and asexual lineages, notably in plant parasitic fungi.

1 Introduction

Sexual reproduction is often associated with ecological functions, such as the formation of structures able to persist, disperse, or both (Williams, 1975; Abrahamson, 1980). Even in species capable of both sexual and asexual reproduction, *i.e.* cyclic parthenogens, sexual reproduction is often the only way to produce structures able to resist winter frost or summer drought. In aphids for instance, eggs are the only cold-resistant stages (Simon et al., 2002). In the cladocerans *Daphnia* (Tessier and Cáceres, 2004) or *Bythotrephes* (Wittmann et al., 2011), diapausing eggs are encapsulated in a cold- and drought-protective envelope. The same applies to cyclically parthenogenetic rotifers (Carmona et al., 2009). Winter survival is also associated with resistance stages in therophyte plants which complete their life cycle during the favorable season and produce seeds to cope with the adverse season. Many plant pathogens, such as fungi and oomycetes, produce thick-walled structures able to cope with adverse environments and to germinate under favorable conditions, *e.g.*, oospores in oomycetes, ascospores in fungi.

Beyond the well-known cost of meiosis (Williams, 1975), sexual reproduction can have demographic consequences. As an example, the lower the population density, the lower the likelihood of finding a mate. Sexual reproduction therefore entails a positive correlation between one component of growth rate (mating success) and population density, called a component Allee effect (Allee et al., 1949). In fact, mating limitation has been reported as the most common mechanism leading to Allee effects (Gascoigne et al., 2009). These effects associated with sexual stages have been detected in several taxa such as the wild radish *Raphanus sativus* L. (Elam et al., 2007), the estuarine grass *Spartina alterniflora* (Davis et al., 2004) and the European gypsy moth *Lymantria dispar* (Vercken et al., 2011). Under some circumstances, such component Allee effects may give rise to extinction threshold population densities, a phenomenon called a strong Allee effect. The occurrence of such an Allee effect depends on the features of reproductive

systems (e.g., overwintering sexual stages, selfing) as well as on social or spatial clustering. Determining whether or not component Allee effects translate into extinction thresholds has triggered numerous studies (Kramer et al., 2009). Yet the consequences of mating limitation have seldom been considered in parasites (Garrett and Bowden, 2002; Krkošek et al., 2012).

Organisms having a dual sexual-asexual reproductive mode are expected to be less vulnerable to Allee effects than exclusively sexual ones. Such a mode of reproduction is typical of many parasites, including aphids, oomycetes and fungi. Interestingly, it has been reported that the level of investment in asexual reproduction is variable within species with obligate sexual reproduction (Simon et al., 2002; Tessier and Cáceres, 2004; Carmona et al., 2009). For instance, European populations of the ascomycete *Leptosphaeria maculans*, which causes stem canker (blackleg) on *Brassica* species, are almost exclusively sexual, while Western Canadian populations invest more in asexual reproduction (Dilmaghani et al., 2012). In Saskatchewan province, polymorphism in sexual mode of reproduction has even been reported. Polymorphism also occurs in poplar rust, caused by the basidiomycete *Melampsora larici-populina*; in France, several genetic groups with markedly different asexual reproduction levels have been reported (Xhaard et al., 2011). In this study, we focus on polymorphism at such a relatively small spatial scale.

Since asexual and sexual reproduction rely on the production of different structures, one may expect a resource allocation dilemma between them (Chamberlain et al., 1997; Schoustra et al., 2010). High investment in asexual reproduction should to some extent lead to a detrimental effect on sexual offspring quantity (Michelmore and Ingram, 1980; Chamberlain et al., 1995) or viability, as repeatedly detected in perennial plants (Vallejo-Marín et al., 2010). A trade-off between asexual reproduction and offspring survival when the host is absent has been reported *in natura* for several plant pathogenic fungi (Carson, 1998; Abang et al., 2006; Sommerhalder et al., 2010, 2011).

Trade-offs between parasite asexual reproduction capacities and survival out of the host have

also been previously considered in studies on parasites that produce free-living forms (Gandon, 1998; Day, 2002; Caraco and Wang, 2008), and has for instance been observed in phages of *Escherichia coli* (De Paepe and Taddei, 2006). Such a trade-off was theoretically shown to support the emergence of polymorphism in strategies of survival outside the host where parasites relying on direct host-to-host transmission coexist with parasites able to survive through free-living forms (Boldin and Kisdi, 2012). A model of evolution of plant pathogens living in seasonal environments showed conditions for coexistence between parasites with very contrasted abilities to overwinter (Hamelin et al., 2011). It also demonstrated that the emergence and maintenance of polymorphism was attributable to negative density-dependence at the onset of the epidemic, *i.e.* during primary infections, since it cannot be explained by the trade-off alone (van den Berg et al., 2011).

In cyclic parthenogens, negative density-dependence during the season (population regulation through competition for resources) alternates with positive density-dependence at the end of the season, *i.e.* component Allee effects associated with sexual reproduction. In other words, population density can be viewed as a common good for mating. Because asexual reproduction trades off with survival, reproducing asexually within the season may be interpreted as a heavy investment in the common good, and investing much in sexual reproduction as a defection. Since cooperation can generate evolutionary diversification (Doebeli and Dieckmann, 2000; Day and Young, 2004; Doebeli et al., 2004; Doebeli, 2011), we wondered whether obligate sexual survival could generate disruptive selection for investment in asexual reproduction, with evolutionary coexistence of sexual (defectors) and asexual (cooperators) lineages.

To investigate this, we model the seasonal dynamics of a pathogen population, using a semi-discrete formalism (Mailleret and Lemesle, 2009). The key feature of the model is that within-season asexual reproduction is followed by an obligate sexual episode at the end of season. Sexual survival forms therefore constitute the primary inoculum for the next season. We focus on

plant parasitic fungi to make explicit model assumptions but keep the model as simple as possible for the sake of generality. First, we analyze population dynamics over ecological timescales, showing that the component Allee effect associated with sexual reproduction translates into an extinction threshold density for the parasite population, *i.e.* a strong demographic Allee effect. Then, we consider the fate of a mutant differing in its investment in sexual reproduction and describe the complex evolutionary outcomes associated with Allee effects. Finally, we discuss the relevance of these results with respect to observed variability in asexual investment.

2 Ecology

2.1 Model

Let (P, S, I) denote the densities of primary inoculum (sexual survival forms), susceptible and infected host, respectively. T denotes the length of one cycle, $\tau < T$ the time during which the host is present, *i.e.*, the growing season length, and n a cycle index. N denotes the total host density, assumed constant during a growing season. Lastly, sexual reproduction is assumed to result from the interaction between individuals with compatible mating types. We also consider only two mating types (designated as $-$ or $+$) as in most species (Billiard et al., 2011). I_- and I_+ denote the densities of hosts infected by $-$ or $+$ mating types, respectively ($I = I_- + I_+$). The following steps of model definition are schematically represented in figure 1.

During host presence. Secondary infection dynamics are described as an SI model (Anderson and May, 1991). New infections occur at a rate proportional to both susceptible host density (S) and infected host density (I). The secondary infection rate β is identical for both mating types.

Let $\dot{I} = dI/dt$. For all t between nT and $nT + \tau$, the model reads

$$\begin{cases} \dot{I}_- = \beta(N - I_- - I_+)I_- , \\ \dot{I}_+ = \beta(N - I_- - I_+)I_+ , \end{cases} \quad (1)$$

where $(N - I_- - I_+) = S$. To keep the model as simple as possible, we have considered that infected hosts remain infectious throughout the season. This simplification fits many plant diseases, such as soil-borne fungal diseases (Bailey and Gilligan, 1999) or viral diseases (Fabre et al., 2012). The resulting dynamics are mathematically equivalent to logistic growth so that the model is not restricted to plant diseases at this stage.

Transition from host presence to host absence. At time $t = nT + \tau$, right before hosts disappear, each infected host produces survival forms following sexual reproduction. In fungal plant pathogens, sexual reproduction may result from contact between two adjacent lesions with compatible mating types, or from fecundation by air-borne gametes as in *e.g.*, *M. larici-populina*. To keep mathematical tractability, the mating function considered is simply bilinear in the densities of + or – mating types:

$$P(nT + \tau) = \Gamma I_- (nT + \tau) I_+ (nT + \tau). \quad (2)$$

During host absence. Survival forms P are subjected to a constant between-season mortality rate μ . Then, for all t between $(nT + \tau)$ and $(n + 1)T$,

$$\dot{P} = -\mu P. \quad (3)$$

One can solve equations (2) and (3), leading to

$$P((n + 1)T) = \Gamma I_- (nT + \tau) I_+ (nT + \tau) e^{-\mu(T - \tau)}. \quad (4)$$

Transition from host absence to host presence. At the beginning of each new season (time $t = (n + 1)T$), a density N of susceptible hosts is made available to the parasite, *e.g.* germinates or is planted. The primary inoculum germinates and infects susceptible hosts, generating ‘–’ or ‘+’ progeny in equal probability. The primary inoculum converts into infected hosts with a factor θ/ξ . This conversion factor derives from a model reduction technique (Appendix A.1) based on the assumption that primary infections occur over a shorter time scale than subsequent plant-to-plant secondary infections (Mailleret et al., 2012). Assuming that θ/ξ is small, the model reads:

$$\begin{cases} I_{-}((n+1)T) &= \frac{1}{2} \frac{\theta}{\xi} P((n+1)T), \\ I_{+}((n+1)T) &= \frac{1}{2} \frac{\theta}{\xi} P((n+1)T). \end{cases} \quad (5)$$

For the sake of simplicity, we stick to this assumption in the remaining of this section. Appendix A.1 presents the model that was actually used in this study; it is slightly more technical but holds for any θ/ξ value.

Thence, the infection cycle (1-4-5) starts again.

Model simplification. Eq. (5) shows that after the first episode of sexual reproduction, the mating-type ratio remains unbiased forever. In addition, model (1-4-5) can be simplified by setting $I = I_{-} + I_{+}$. Let $\chi = (\theta/\xi)(\Gamma/4)e^{-\mu(T-\tau)}$. The model now reads, for all t between nT and $nT + \tau$,

$$\dot{I} = \beta(N - I)I, \quad (6)$$

and

$$I((n+1)T) = \chi I(nT + \tau)^2. \quad (7)$$

Model (6-7) can be further reduced to:

$$I((n+1)T) = \chi \left(\frac{I(nT)N}{Ne^{-\beta N\tau} + I(nT)(1 - e^{-\beta N\tau})} \right)^2. \quad (8)$$

The above discrete-time equation maps the density at the beginning of one season $I(nT)$ onto the density at the beginning of the next season $I((n+1)T)$.

2.2 Strong Allee effect

At low infected host density ($I \ll N$), Eq. (8) simplifies to

$$\frac{I((n+1)T)}{I(nT)} \approx \chi e^{2\beta N\tau} I(nT), \quad (9)$$

which shows that parasite reproductive ratio increases with infected host density *i.e.*, the parasite population undergoes a component Allee effect. This is a direct consequence of the mating function used in Eq. (2).

The reproductive ratio $I((n+1)T)/I(nT)$, computed from Eq. (8) as a function of population density $I(nT)$, determines the fate of the parasite population (figure 2; Appendix A.2). It may never reach 1 (figure 2A-C), so that the dynamics converge to the disease free equilibrium ($I = 0$) and the parasite goes extinct whatever its initial density. If parameter values are such that the reproductive ratio reaches 1 for certain densities (figure 2B-D), Eq. (8) has three fixed points: the disease free equilibrium and an endemic equilibrium I_e , which are both stable, as well as a critical unstable fixed point I_c ($0 < I_c < I_e$), separating the dynamics between extinction ($I < I_c$) and convergence to the endemic equilibrium I_e ($I > I_c$). If the reproductive capacities of the parasite are high enough, and provided its initial population density is large enough, it may persist on a year-to-year equilibrium, generating T -periodic epidemic outbreaks. The model cannot show

year-to-year cycles or chaotic behavior when a single morph is considered.

3 Evolution

We will now turn to investigating the evolutionary implications of these Allee effects. Following an adaptive dynamics approach (Metz et al., 1992; Dieckmann and Law, 1996; Dieckmann, 2004), we have derived an invasion criterion for an initially rare mutant phenotype, challenging a distinct resident phenotype.

The secondary infection rate β represents investment in asexual reproduction, as it is proportional to the number of asexual spores produced per infected individual per unit time within the season. We assumed that β trades off with survival ability. The *per capita* mortality rate μ is therefore an increasing function f of β . In what follows, we refer to the pair (β, μ) as a single ecological trait which is further assumed to depend on a single multi-allelic locus. Mating is also assumed to be coded by a single independent locus. We have focused on a haploid species in which survival forms result from meiosis (Heitman et al., 2007; Coluccio et al., 2008). As a consequence, sexual survival forms inherit each allele from either parental cell with equal probability for both the mating type and the ecological trait.

3.1 Evolutionary invasion analysis

We have assumed that evolution proceeds through successive mutations arising within a resident population which have reached permanent ecological dynamics, as opposed to transient ones. The dynamics of each emerging mutant can be predicted by deriving an invasion criterion. To this end, we have extended model (6-7) to two subpopulations (resident and mutant), denoted by subscripts 1 and 2.

Each subpopulation contains both mating-types ‘−’ and ‘+’ in equal proportions (Section 2.1).

For instance, $I_1 = I_{1,+} + I_{1,-}$ represents the subpopulation having the ecological trait (β_1, μ_1) .

During the season (t between nT and $nT + \tau$), the continuous part is described by:

$$\begin{cases} \dot{I}_1 &= \beta_1 I_1 (N - I_1 - I_2) , \\ \dot{I}_2 &= \beta_2 I_2 (N - I_1 - I_2) . \end{cases} \quad (10)$$

At the end of the season, offspring is generated from the mating of two parents that have either the same ecological trait, as in Eq. (7), or two different traits, *i.e.* $I_{1,-} \times I_{2,+} + I_{1,+} \times I_{2,-} = 2 \times (I_1/2)(I_2/2)$. In the latter case, because the parasite is haploid, each offspring has probability $1/2$ of inheriting either parental allele for each trait. Therefore, offspring production for each trait, at time $t = nT + \tau$, is

$$\begin{cases} P_1(t) &= \Gamma \left(\underbrace{\left(\frac{I_1(t)}{2} \right)^2}_{\text{resident} \times \text{resident}} + \frac{1}{2} \times 2 \times \underbrace{\frac{I_1(t)}{2} \frac{I_2(t)}{2}}_{\text{resident} \times \text{mutant}} \right) , \\ P_2(t) &= \Gamma \left(\underbrace{\left(\frac{I_2(t)}{2} \right)^2}_{\text{mutant} \times \text{mutant}} + \frac{1}{2} \times 2 \times \underbrace{\frac{I_1(t)}{2} \frac{I_2(t)}{2}}_{\text{resident} \times \text{mutant}} \right) . \end{cases} \quad (11)$$

Taking into account inter-season survival and primary infections results in the following discrete part:

$$\begin{cases} I_1((n+1)T) &= \chi_1 (I_1(nT + \tau)^2 + I_1(nT + \tau)I_2(nT + \tau)) , \\ I_2((n+1)T) &= \chi_2 (I_2(nT + \tau)^2 + I_1(nT + \tau)I_2(nT + \tau)) . \end{cases} \quad (12)$$

where $\chi_i = (\theta/\xi)(\Gamma/4)e^{-\mu_i(T-\tau)}$ and $\mu_i = f(\beta_i)$, for $i = 1, 2$.

When no mutant is present, the dynamics of a resident population with trait β_1 are governed by Eq. (6-7) and therefore converge to a T -periodic invariant solution, $I_1^\circ(t, \beta_1)$ (Section 2.2). Since the mutant is rare initially ($I_2 \ll I_1$), its effects on the resident dynamics are negligible at

first. During the first season

$$\begin{aligned} \dot{I}_2 &\approx \beta_2 I_2 (N - I_1^\circ), \\ I_2(T) &\approx \chi_2 I_1^\circ(\tau, \beta_1) I_2(\tau). \end{aligned}$$

Let $\bar{I}_1^\circ = (1/\tau) \int_0^\tau I_1^\circ(t, \beta_1) dt$, so that $N - \bar{I}_1^\circ$ denotes the mean susceptible host density at mutant-free equilibrium. We thus have $I_2(\tau) \approx I_2(0) e^{\beta_2 \tau (N - \bar{I}_1^\circ)}$. The mutant can invade provided that $I_2(T) > I_2(0)$, which reads

$$\frac{I_2(T)}{I_2(0)} \approx I_1^\circ(\tau, \beta_1) \times \chi_2 e^{\beta_2 \tau (N - \bar{I}_1^\circ)} > 1, \quad (13)$$

whereas if this quantity is less than one, the mutant goes extinct.

It is worth noting that the above invasion criterion is determined by two variables shaped by the resident trait, *i.e.* mean susceptible host density ($N - \bar{I}_1^\circ$) and infected host density at the end of the season $I_1^\circ(\tau, \beta_1)$. Moreover, there is a positive relationship between the end-of-season resident population density and the mutant reproductive success. Because the mutant is initially rare, it mainly mates with the resident. As a consequence, higher end-of-season resident population density results in more frequent mutant sexual reproduction which enables it to overwinter.

3.2 Evolutionary outcomes

By definition of the endemic equilibrium, one has:

$$\frac{I_1^\circ(T)}{I_1^\circ(0)} = I_1^\circ(\tau, \beta_1) \times \chi_1 e^{\beta_1 \tau (N - \bar{I}_1^\circ)} = 1. \quad (14)$$

Therefore, the invasion criterion (13) reads

$$\frac{I_2(T)}{I_2(0)} \approx \frac{\chi_2 e^{\beta_2 \tau (N - \bar{I}_1^\circ)}}{\chi_1 e^{\beta_1 \tau (N - \bar{I}_1^\circ)}} > 1. \quad (15)$$

We can thus derive an invasion fitness function as

$$s(\beta_1, \beta_2) = (\beta_2 - \beta_1)(N - \bar{I}_1^\circ(\beta_1))\tau - (f(\beta_2) - f(\beta_1))(T - \tau). \quad (16)$$

The invasion fitness $s(\beta_1, \beta_2)$ is positive if phenotype β_2 can invade resident phenotype β_1 ; it is negative otherwise. The (β_1, β_2) pairs for which the mutant initially invades or not are illustrated in pairwise invasibility plots (PIP, figures 3A–6A).

Under the trade-off hypothesis, there may exist ranges of β trait values where ecological dynamics of a monomorphic population converge to the disease-free equilibrium which is actually the only stable equilibrium (illustrated with light gray zones in figures 3B–6B). These phenotypic ranges are henceforth referred to as a monomorphic extinction zones.

A necessary condition for a monomorphic population to split into a dimorphic population is given by

$$\left. \frac{\partial^2}{\partial \beta_2^2} s(\beta_1, \beta_2) \right|_{\beta_1 = \beta_2 = \beta} = -f''(\beta)(T - \tau) > 0. \quad (17)$$

Hence, evolutionary branching is conditioned by the concavity of the trade-off function: it can only occur under concave trade-off shapes ($f'' < 0$).

Numerical simulations performed for model (10-12), using the algorithm detailed in Appendix B.2 and implemented in MATLAB version R2012b (Supplementary Material), showed that under convex trade-offs, any initial population with β value above the monomorphic extinction zone continuously evolves towards a singular trait value β^* . Provided reaching β^* does not require entering an extinction zone, the population monomorphically converges and stabi-

lizes to β^* trait value (figure 3C,D). By contrast, whenever reaching β^* requires entering into a monomorphic extinction zone, the population goes extinct (figure 4C,D). The density of the mutant increases at first, taking advantage of the presence of the resident. During this phase, the resident population decreases to extinction. The mutant eventually predominates but it also goes extinct as it is unable to survive alone.

A concave trade-off is a necessary yet not sufficient condition for an initially monomorphic population to evolve towards a branching point. When it does, as predicted by the PIP from figure 5A, the parasite population splits into a morph investing decreasingly into asexual reproduction (morph (a)), and a morph doing the opposite (morph (b)). This dimorphism allows morph (a) to cross the monomorphic extinction zone, because morph (b) provides it with mating opportunities. Morph (a) then becomes exclusively sexual ($\beta_{min} = 0$). The higher density corresponds to the ecological dynamics of an exclusively sexual morph (a), with a cyclic parthenogen (b) occurring at lower densities. For large β_{max} values, the cyclic parthenogen can also go extinct, and the population becomes exclusively sexual (figure 6B). The relative increase in morph (b) density benefits the exclusively sexual morph (a), which ends up saturating the environment and excluding morph (b). Such unilateral extinction is made possible by the fact that morph (a) reached the critical population density to be able to survive alone ($I_a > I_c$ for $\beta = 0$, see figures 6B and 7).

4 Discussion

4.1 Strong Allee effect and evolutionary extinction

That the component Allee effect linked to obligate mating causes a strong demographic Allee effect in cyclic parthenogens is an important conclusion of this work. This may lead the population to evolve to trait values where persistence is no longer possible, a phenomenon that has

been termed “evolutionary extinction” or “evolutionary suicide” (Webb, 2003; Parvinen, 2005). Whether or not these critical parameter ranges can be reached in nature is not yet known. One may imagine that these ecological and evolutionary consequences of mate limitation might ultimately explain why many species in temperate areas do not undergo obligate sexual reproduction at the end of the season but rather adopt a mixed sexual-asexual survival strategy (Barrett et al., 2008).

4.2 A novel mechanism promoting variability in reproductive strategies

Allee effects, *i.e.* positive density dependence, can promote evolutionary diversification as well. Indeed, we showed that obligate sexual reproduction makes the invasion fitness depend on both the mean resident density and the resident end-of-season density, which act on mutant invasion fitness in opposite ways. The mean resident density decreases mutant fitness, since both compete for susceptible hosts; the end-of-season resident density increases mutant fitness because it enables the mutant to generate survival forms through sexual reproduction.

Under convex transmission-survival trade-offs, these opposite effects select for a mixed sexual-asexual reproduction in cyclic parthenogens. The corresponding intermediate investment into asexual reproduction has to be large enough to overcome Allee effects but small enough to ensure sexual survival. By contrast, under concave trade-offs, selection is disruptive and the population may undergo evolutionary branching.

After evolutionary branching occurs, continual phenotypic divergence is observed. Thus, the population comprises a cyclical parthenogenetic morph with both high asexual reproduction rate and low survival ability and an exclusively sexual morph, with high survival. When maintained, such a dimorphism is a “tragedy of the commune” (Doebeli et al., 2004). Evolution leads to an asymmetric state in which asexual morphs make cooperative investments in population density

(a common good given the strong Allee effect) at the expense of a lower survival, while sexual morphs invest nothing, survive better, and eventually reach higher densities.

4.3 Transitions from cyclical parthenogenesis to strict sexual reproduction

Evolution may alternatively drive the asexual morph to extinction. Although the sexual morph has maximal survival capacities, it cannot survive as a single morph before it has reached a quite elevated critical Allee density. During the transition, the sexual morph benefits from the presence of the asexual morph as it increases sexual reproduction opportunities. Once the sexual morph has grown above the Allee threshold, it may eventually exclude the asexual morph. Such a phenomenon is termed “evolutionary murder” by some authors (Parvinen, 2005). Asexual reproduction can thus be an evolutionary transient state leading to strict sexual reproduction. A similar phenomenon is highlighted by Boldin and Kisdi (2012), who show that evolutionary branching can ultimately select for a single parasite strain which excels in transmission via free-living forms. An open-ended question is whether such complex evolutionary transitions might happen in nature. Phylogenetic reconstructions of character states may help with this issue, provided that adaptation is a slow process, long-lasting enough to be detected by phylogenetic methods.

4.4 Disentangling ultimate causes of reproductive polymorphism

Possible mechanisms for polymorphism in sexual mode of reproduction include (i) temporal variation in winter survival rates (Simon et al., 2002), (ii) local adaptation to fine-scale spatial variation (Kawecki and Ebert, 2004), and (iii) Allee effects associated with obligate sexual reproduction, as shown in this study.

From their analysis of *L. maculans* populations, Dilmaghani et al. (2012) suggest that poly-

morphism in sexual mode of reproduction may be related to climatic conditions in Saskatchewan province, which is akin to (i). In *M. larici-populina*, Xhaard et al. (2011) suggest that sexual and almost asexual lineages evolved at the ends of an altitudinal gradient and happen to co-occur at intermediate positions in the gradient, which is akin to (ii). Alternatively, it may be that in both species, polymorphism results from (iii), Allee effects and obligate sexual reproduction.

Distinguishing between adaptive hypotheses can be challenging. Testing adaptation to temporal variability in winter survival rates requires relating the variance in reproductive strategies to climatic data as exemplified by Halkett et al. (2004) in aphids. Testing whether polymorphism in reproductive strategies is selected as a local adaptation in a spatially heterogeneous environment implies reciprocal transplantations (Lenormand, 2002). If this mechanism is not involved, sexual and almost asexual lineages should perform as well in every tested environment. Testing whether Allee effects explain variability could be done by experimentally manipulating morph frequencies, such as testing an advantage of the rarest in dimorphic plants (Gigord et al., 2001).

Lastly, an alternative mechanism for polymorphism involves multiple infections, as shown by Alizon and van Baalen (2008) and Boldin and Diekmann (2008). Given a reasonably sensitive genetic marker, it might be feasible to test this hypothesis (López-Villavicencio et al., 2011).

4.5 Perspectives

From a broader perspective, our study may shed light on the coexistence of male and non-male producing clones in cyclic parthenogens such as *Daphnia* (Galimov et al., 2011). Producing males is likely a costly venture. However, non-male producing morphs still have to mate with male producing morphs to complete their life cycle. In our study, strictly sexual morphs similarly depend on cyclic parthenogens for obligate mating to take place. By the same token, non-male producing morphs (defectors) may indeed coexist with male-producing morphs (cooperators).

Since we considered sexual reproduction obligate, our model has not been designed to analyze whether sexual reproduction can out-compete asexual reproduction on an evolutionary timescale. Nevertheless, the mechanism involved in the loss of a dual reproductive mode towards exclusive sexual reproduction highlights the fact that complex dynamics may lead cyclic parthenogens to become exclusively sexual. Further studies may reconsider the hypothesis of an exclusively sexual overwintering survival, and investigate asexual survival possible as well. This would pave the way toward addressing the emergence and the coexistence of exclusively sexual and exclusively asexual lineages in nature, as well as the possible loss of sex.

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A Ecological model

A.1 Derivation of Eq. (5)

Here we briefly present the aggregation technique proposed by Mailleret et al. (2012) to reduce semi-discrete seasonal epidemic models. This technique is based on the hypothesis that primary infections occur on a faster time scale than secondary infections. A full in-season model

accounting for primary infections, primary inoculum depletion, and secondary infections reads:

$$\begin{aligned}\dot{P} &= -\frac{\xi}{\varepsilon}SP, \\ \dot{S} &= -\frac{\theta}{\varepsilon}SP - \beta S(I_- + I_+), \\ \dot{I}_- &= \frac{1}{2}\frac{\theta}{\varepsilon}SP + \beta SI_-, \\ \dot{I}_+ &= \frac{1}{2}\frac{\theta}{\varepsilon}SP + \beta SI_+, \end{aligned}$$

where each primary infection generates ‘−’ or ‘+’ infected hosts in equal probability. Parameters ξ and θ are associated with primary inoculum depletion and primary infections, and ε is a scaling factor. At the beginning of the season $S((n+1)T) = N$, $I_+((n+1)T) = 0$, $I_-((n+1)T) = 0$ and $P((n+1)T) > 0$.

Assuming that primary infections occur on a faster time-scale than secondary infections amounts to considering that ε is small. Rewriting the model in the time-scale corresponding to primary infections ($t' = t/\varepsilon$), and neglecting the terms in ε , the model reads:

$$\begin{aligned}P' &= -\xi SP, \\ S' &= -\theta SP, \\ I' &= \theta SP, \end{aligned}$$

with the prime indicating derivatives with respect to t' and $I = I_+ + I_-$. Thus, on the time-scale of primary infections, the quantity $S - (\theta/\xi)P$ remains equal to $N - (\theta/\xi)P((n+1)T)$, so that one has:

$$S' = -\xi S \left(S - \left(N - \frac{\theta}{\xi} P((n+1)T) \right) \right).$$

Since $S((n+1)T) = N > 0$, S converges to $\max(0, N - (\theta/\xi)P((n+1)T))$ over the time scale of primary infections. Accordingly, since $S + I = N$, I converges to $\min((\theta/\xi)P((n+1)T), N)$ over the time-scale of primary infections. When the primary inoculum is very infectious (large

θ/ξ), it can infect every susceptible hosts during the primary infection phase. When θ/ξ is small, primary infections result in the linear conversion of primary inoculum into infected host, as in Eq. (5), resulting in model (8).

A.2 Equilibria

Model (8)'s equilibria are fixed points $I^* = m(I^*)$ of the discrete-time map $m : I_n \mapsto I_{n+1}$ from Eq. (8). The disease-free equilibrium (DFE) $I = 0$ is a trivial fixed point. Apart from the DFE, other equilibria may exist, depending on the following conditions.

If $\beta = 0$, one finds one additional fixed point: $I = 1/\chi$. Otherwise ($\beta > 0$), two additional fixed-point may exist, provided that

$$\Delta = \chi N(\chi N - 4e^{-\beta N\tau}(1 - e^{-\beta N\tau})) > 0. \quad (18)$$

Let

$$\begin{aligned} I_c &= \frac{N}{2} \frac{\chi N - 2e^{-\beta N\tau}(1 - e^{-\beta N\tau}) - \sqrt{\Delta}}{(e^{-\beta N\tau} - 1)^2}, \\ I_e &= \frac{N}{2} \frac{\chi N - 2e^{-\beta N\tau}(1 - e^{-\beta N\tau}) + \sqrt{\Delta}}{(e^{-\beta N\tau} - 1)^2}, \end{aligned} \quad (19)$$

thereafter referred to as critical and endemic equilibrium densities, respectively. One can easily check that $\Delta > 0$ implies that $I_e > I_c > 0$.

An equilibrium I^* is stable if $|m'(I^*)| < 1$ and unstable if $|m'(I^*)| > 1$. Since $m'(0) = 0$, the DFE ($f(0) = 0$) is stable in any event. Similarly, one can notice that for $\beta = 0$, $m'(1/\chi) = 2$ so that this non generic fixed-point is always unstable. Moreover, one can easily check that m is a continuous and strictly increasing function, so that a graphical analysis suffices to show that I_c and I_e are unstable and stable, respectively (Figure 2).

B Phenotypic competition and evolution

B.1 Competition model

Considering a set of phenotypes $\{\beta_i\}$, $i = 1, 2, \dots, n$, the in-season competition model reads:

$$\begin{aligned}\dot{P}_i &= -\frac{\xi}{\varepsilon}SP_i, \\ \dot{S} &= -\sum_j \frac{\theta}{\varepsilon}SP_j - \sum_j \beta_j SI_j, \\ \dot{I}_i &= \frac{\theta}{\varepsilon}SP_i + \beta_i SI_i.\end{aligned}$$

Proceeding as in Appendix A.1, and noticing that $r_i = P_i / \sum_j P_j$ is a constant, we obtain, during the season,

$$\begin{aligned}\dot{S} &= -\sum_j \beta_j SI_j, \\ \dot{I}_i &= +\beta_i SI_i,\end{aligned}\tag{20}$$

and from season to season:

$$\begin{aligned}S((n+1)T) &= \max\left(N - \frac{\theta}{\xi} \sum_j P_j((n+1)T), 0\right), \\ I_i((n+1)T) &= \min\left(\frac{\theta}{\xi} P_i((n+1)T), N \times \frac{P_i((n+1)T)}{\sum_j P_j((n+1)T)}\right),\end{aligned}\tag{21}$$

where

$$P_i((n+1)T) = \frac{\Gamma}{4} e^{-\mu_i(T-\tau)} \left(I_i(nT + \tau) \sum_j I_j(nT + \tau) \right),\tag{22}$$

see equation (11). When θ/ξ is small, primary infections results in the linear conversion of primary inoculum into infected hosts, with I_i tending to $(\theta/\xi)P_i((n+1)T)$, which leads to Eq. (12).

Evolutionary computations were actually realized using the full model (20-21-22).

B.2 Evolution algorithm

The evolving phenotype β ranges from 0 to β_{\max} . This interval is divided into a finite number of subintervals ($n = 25$). Simulations start from a monomorphic population having a given β trait value. Once at ecological attractor, which we compute as the solution of the full competition model after a fixed horizon of 10,000 years, a small mutation occurs, having an equal probability of being on the left (smaller β) or on the right (larger β), regardless the subinterval considered. Then the process reiterates.

B.3 Evolutionary exclusion dynamics

We additionally provide figure 7.

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Figures legends

Figure 1: Schematic representation of the different epidemic processes in the seasonal plant epidemic model. P (the primary inoculum) denotes survival forms resulting from sexual reproduction of two individuals having compatible mating types. S and I denote susceptible and infected individual densities, respectively. Dashed lines represent processes occurring at the beginning of the season (plants renewal and primary infections), solid lines represent secondary infections occurring during a growing season, dashed-dotted lines represent sexual reproduction occurring at the end of the host growing season, and the dotted line represents mortality of the sexual forms between growing seasons.

Figure 2: Graphical illustration of the year-to-year model (8) in the two generic cases: global extinction (left panels) and bi-stability typical of a strong Allee effect (right panels). Top row: year n to year $(n + 1)$ reproductive ratio as a function of infected host density at the beginning of year n . Bottom row: cobweb diagrams plotting infected host density at the beginning of year $(n + 1)$ as a function of that at the beginning of year n . Gray squared-lines are typical season-to-season dynamics produced by model (8) across years, eventually converging to the disease free equilibrium (panel C and panel D) or to the endemic equilibrium I_e (panel D).

Figure 3: Evolutionary process leading to a monomorphic evolutionary endpoint (Section 3.2). (A.) Pairwise invasibility plot (PIP). Areas where the mutant can invade are in dark gray. (B.) Equilibria in the mutant-free ecological model, as a function of the resident trait: disease-free equilibrium ($I = 0$, stable), I_c (unstable) and I_e (stable) or saturation ($I = N$, stable); black and gray colors correspond to stable and unstable equilibria, resp. (C.) The associated evolutionary dynamics; numerical simulations were realized using an algorithm which is described in Appendix B.2. The light gray areas (in panels A., B. and C.) correspond to resident trait values for which the endemic equilibrium I_e does not exist *i.e.*, to a monomorphic extinction zone (Appendices B and A.2; Panel B.). (D.) Ecological dynamics at evolutionary endpoint. For all panels, we used the following trade-off form: $\mu = f(\beta) = c\beta^a$, here $a = 1.2$ (trade-off shape: convex) and $c = 0.25$ (trade-off coefficient), $\beta_{min} = 0$ and $\beta_{max} = 6$ (phenotypic boundaries), $T = 1.5$ (time unit), $\tau = 1$ (time unit), $N = 1$ (host unit), $\theta = 1.06$ (per primary inoculum unit per unit time), $\xi = 1$ (per host unit per unit time), $\Gamma = 4$ (per squared host unit per unit time).

Figure 4: Evolutionary process leading to extinction. We refer the reader to figure 3's caption since the only change concerns the trade-off shape parameter: here, $a = 2.5$ (more convex trade-off). Also, we took $\beta_{max} = 3$ for illustrative purposes. Last, Panel (D.) shows ecological dynamics *right before* the evolutionary endpoint (extinction) *i.e.*, from the last mutation preceding extinction, and over a far greater number of seasons than in figure 3D.

Figure 5: Evolutionary branching process leading to a dimorphic evolutionary endpoint (Section 3.2). We refer the reader to figure 3 caption since the only change concerns the trade-off shape parameter: here, $a = 0.9$ (concave trade-off).

Figure 6: Evolutionary process leading to extinction of one branch. We refer the reader to figure 3 caption since the only changes concern the trade-off parameters: here, $a = 0.9$ (concave trade-off) and $c = 0.21$ (this deviation from figure 5 is for illustrative purposes only). However, we took $\beta_{max} = 10$ (as compared to 6 in figure 5) to get extinction. Last, Panel (D.) shows ecological dynamics right before the evolutionary endpoint (extinction of one branch) over a far greater number of seasons, and over a far greater number of seasons than in figure 3D.

Figure 7: Zoom on exclusion dynamics in figure 6. (A.) Ecological equilibria for a mutant-free monomorphic population as a function of the parasite trait β : the disease-free equilibrium $I = 0$ and the endemic equilibria $I = I_e$ and $I = N$ (saturation) are stable (black lines), while $I = I_c$ is unstable (gray line). The light gray area corresponds to a monomorphic extinction zone (Appendices B and A.2). (B.) Ecological dynamics leading to morph (b) exclusion. The black and gray lines correspond to morph (a) and (b) dynamics, resp. The darker gray line corresponds to the mutant arising from morph (b). We refer the reader to figure 6 caption for parameter values.